98/8 Doc IIIA section No.	7.1.1.2.1 <i>I</i> 01	Ready biodegradability
91/414 Annex	П	Ready biodegradability
Point addressed	7.2.1.3.1 / 01	

1. Annex point(s) IIA, 7.2.1.3.1 Ready biodegradability

Location in Dossier Section 5
 Authors (year) R.Grade (1996)

Title Report on the test for ready biodegradability of CGA 293343 tech. in the carbondioxide

evolution test

Report No., Date Study No. 95G001, 8. January, 1996.

Syngenta File No(Desire) 293343/31

Owner Syngenta Crop Protection AG

4. Testing facility Ciba Geigy Ltd, Product Safety / Ecotoxicology, 4002 Basel, Switzerland

5. Dates of work September 20, 1995 to October 25, 1995

6. Test substance CGA 293343

**7. Test method** The study was be conducted in compliance with:

OECD Guideline No.: 301/B (Paris 17/07/92) 92/69/EEC C.4-C.

**8. Deviations** Only one CO<sub>2</sub> scrubber was used.

9. GLP This study was performed in compliance with Good Laboratory Practice (GLP) in

Switzerland, Procedures and Principles, March 1986 [Verfahren und Grundsaetze der Guten Laborpraxis (GLP) in der Schweiz, Maerz 1986] issued by the Federal Department of the Interior and the Intercantonal Office for the Control of Medicaments, Switzerland. These procedures are based on OECD Principles of GLP adopted on 12 May 1981 by Decision of the OECD Council concerning Mutual Acceptance of Data in the Assessment

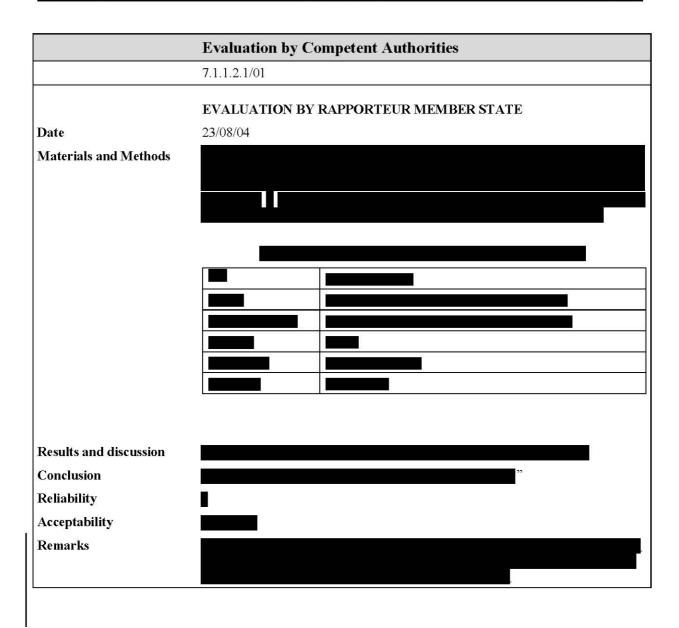
of Chemicals [C(81)30 (Final)].

Test system: The test was performed with thiamethoxam technical grade of 98.6% purity. For the toxicity control, where the test substance and reference were applied together, the reference amount was 15.0 mg DOC/l and the test substance amount was 15.6 mg TOC/l. During exposure the evolved CO<sub>2</sub> was measured at 0, 3, 6, 8, 10, 13, 17, 21, 24, 28 and 29 days. The test system is described Table 1.

Table 1: Test system for carbon dioxide evolution study (Grade 1996)

pH:	8.0 (after collection)
Source:	Sewage treatment plant, CH-4153 Reinach, Switzerland.
Test concentration:	45.0 - 47.9 mg /l, corresponding to 14.8 - 15.8 mg TOC/l
Duration:	29 days
Temperature	$22 \pm 2$ °C, diffuse light
Test water:	Distilled water

Findings: The biodegradation of the test substance was 7% in 29 days, therefore the test substance was not readily biodegradable in this test. thiamethoxam did not inhibit the biodegradation of the reference substance (sodium benzoate). According to the 7th Amendment to Directive 67/548/EEC, i.e. Directive 92/32/EEC, the ecotoxicological classification for thiamethoxam is: "not readily biodegradable".



98/8 Doc I section No.	IIA 7.1.2.1.2	Anaerobic biodegradation
91/414 Ann Point addressed	7777	Supplementary soil degradation studies - anaerobic degradation

1.2 Title Anaerobic Aquatic Metabolism of (14C-thiazole) CGA 293343

Report and/or project N° Syngenta<sup>o</sup>(Desire)

ABR-96099 293343/468

1.4 Lab. Report Nº

507-95

Location in Dossier

Section 5.

1.6 Authors

1.7

Adora Clark, Ph.D. March 13, 1998

Date of report 1.8 Published / owner

Unpublished report, Syngenta Crop Protection AG

2.1 **Testing facility** 

Novartis Crop Protection Inc Environmental Safety Department, Greensboro, NC 27419,

USA

PTRL West, Incorporated, Richmond, CA 94806, USA

2.2. Dates of work

Study Initiation: 25 September, 1995 Experimental Start: 14 November, 1995

Experimental Termination:

Study Completion: 13 March, 1998

Objectives 3.

Determine formation and decline of degradates from incubation of CGA-293343 under

anaerobic aquatic conditions.

4.1 Test substance ISO common name: thiamethoxam Trade name: Not applicable

Ratch:

<sup>14</sup>C-labeled test substance Specific activity of [.....]

Radiochemical purity of the test

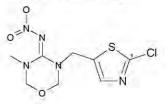
substance:

Structural formula: (\*position of label)

Yes [x]

(14C-thiazole) CGA 293343 1.49 Mbq/mg

labeled in thiazole ring



4.2 Specification See Section 4.1

Storage stability Stable Stability in vehicle Stable 4.4

Homogeneity in vehicle 4.5 Available, but not reported 4.6 Validity Available, but not reported

Vehicle / solvent Acetonitrile 4.7

4.8 Physical form Available, but not reported

5.1 Test method This Anaerobic Aquatic Soil Metabolism study was designed to meet EPA Guidelines for

degradation in aquatic sediment under anaerobic conditions (guideline 162-3).

5.2 Justification Usually, field soils are used as aquatic sediments. Therefore, this design meets essentially

European guidelines as required in: Commission Directive 95/36/EC of 14 July 1995 amending Council Directive 91/414/EEC: Annex I: 7.1.1.1.2. Supplementary Studies:

Anaerobic Degradation

Copy of method Test method available, but not reported

Choice of method Available, but not reported

7 **Deviations** The clay minerology was not conducted under GLP

**Certified laboratory** Not applicable 8.1 Certifying authority Not applicable

8.3 GLP yes

8.4 Justification This study was designed to meet EPA Guidelines for degradation in aquatic sediment

under anaerobic conditions (guideline 162-3).

GEP Not applicable 9.1 Not applicable

9.2 Type of facility (official or officially recognised)

Justification Not applicable

Test system

Water characteristics	Origin	HPLC Grade water purchased from Fisher Scientific, Pittsburgh, PA.
	The water was sterile overnight.	filtered and sparged with nitrogen

Sediment characteristics	Origin	Novartis Crop Protection western research Station, Fresno County, California
	рН	7.3
	Redox potential (mV)	not reported
	Organic matter (%)	0.6
	Total nitrogen (%)	0.056
	Total phosphorus (%)	not reported
	Cation exchange capacity (meq/100 g sediment)	7.4
	Particle size distribution - % clay - % silt - % sand	8 25 67
	Classification (USDA)	Sandy loam
soil/water slurry: Total Anaerobic Bacteria in (CFU/g dry sediment)	Experimental start: Day 90 Day 182 Day 365	0.22 0.10 0.20 0.51

Equilibrium of test system and treatment	
Incubation conditions	Equilibration under nitrogen
Weight of water (mL)	100
Weight of sediment (dry weight g)	50
Temperature	25.0 °C
Exclusion of light	Yes[x] No[]
Equilibrium time of test system	35 days
Treatment rate	0.1 μg/g
Sampling intervals	0, 7, 15, 21, 29, 43, 62, 90, 120, 180, 272, 365 days
Replicates	Yes[x] No[]

11 Statistics kinetics
12 References (published) None
13 Unpublished data None

Test system: The route and rate of degradation of <sup>14</sup>C-thiamethoxam was investigated over 365 days at 2 concentrations: 0.1 mg/kg ('kinetic', chosen based on an application rate of 200g as/ha, 15 cm soil layer, bulk density 1.5) and 5.3 mg/kg ('bulk', 50 x exaggerated application rate) in a sandy loam soil flooded with water and equilibrated under nitrogen. The degradation products were characterised by TLC and HPLC co-chromatography with reference standards and identified by mass-spectroscopy. Further details of the soil characteristics are presented in Table 1 (same soil as used in aerobic experiment, microbial biomass measurements only relevant for aerobic incubation).

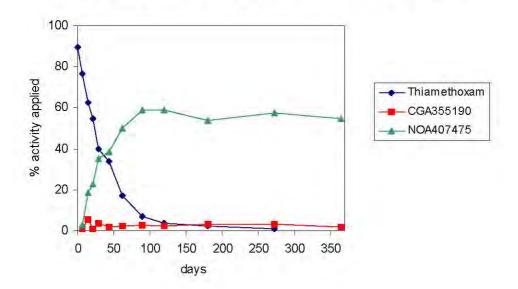
Findings: The mean total applied radioactivity recovered from the sample matrices (material balance) ranged from 90.7% to 98.8% for the kinetic samples and from 89.5% to 95.6% for the bulk samples. Thiazolyl-¹4C labelled thiamethoxam was metabolised quite rapidly under anaerobic conditions to form the denitro compound NOA 407475 as a major metabolite with a plateau value of 58.8% to 58.9% from day 90 to day 120 and a decrease to 54.7% at day 365 (kinetic experiment). For the bulk experiment the peak was reached at day 180 with 57.6% with a decrease to 51.0% at day 365. The hydrolysis product CGA 355190 was also present peaking at 5.4% (kinetic, day 15) and at 18.1% (bulk, day 90) before its decline to 2.0% and 2.1%, respectively. All remaining components were below 2.8% (kinetic) and below 5.2% (bulk) of the dose. Some of the metabolites characterised by co-chromatography were CGA 322704, CGA 353968 and NOA 404617. Volatiles reached a maximum of 2.7% (kinetic) and 4.4% (bulk) of applied radioactivity and were identified as carbon dioxide. Non - extractables reached 19.5% (kinetic) and 13.0% (bulk).

Table 1: Distribution of radioactivity after application of <sup>14</sup>C-Thiazole-thiamethoxam in sandy loam soil, anaerobic conditions, as percent of applied dose (Clark, 1998a).

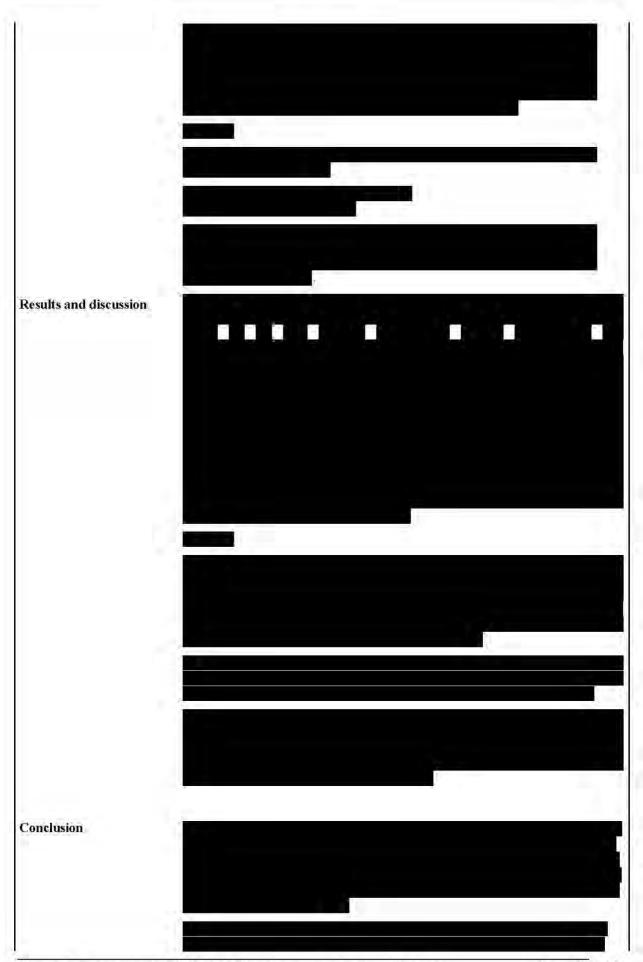
Incubation time [days]	Thiamethoxam [%]	CGA355190 [%]	NOA407475 [%]	Volatiles [%]	non extractables [%]
0	89.2	< LOD	< LOD	< LOD	0.5
7	76.2	1.0	3.1	0.5	4.0
15	62.4	5.5	18.4	0.3	3.1
21	54.9	1.9	22.9	0.3	5.6
29	39.6	3.9	35.4	1.1	5.3
43	34.6	1.8	42.3	0.1	6.0
62	17.0	2.1	50.2	0.2	15.8
90	7,1	3.1	58.9	0.7	18.0
120	3.9	2.5	58.9	1.0	17.2
180	2.6	3.4	53.9	2.7	18.9
272	0.9	3.3	57.3	2.5	19.5
365	< LOD	2.0	54.7	1.6	18.8

LOD: Level of detection

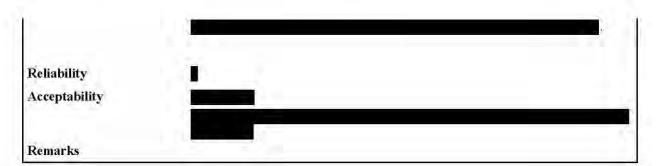
Figure 1: Formation and decline of major metabolites for <sup>14</sup>C- Thiazole -thiamethoxam in a sandy loam soil under anaerobic conditions (Clark, 1998a)



	Evaluation by Competent Authorities	
	Use separate "evaluation boxes" to provide transparency as to the comments and views submitted	
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Materials and Methods		



Section 7: Ecotoxicological Profile Including Environmental Fate and Behaviour



98/8 section	Doc III No.	A 7.1.2.1.2	Anaerobic biodegradation
91/414	Anne	x II	Supplementary soil degradation studies - anaerobic degradation
Point a	ddressed	7.1.1.1.2	

1.	Annex point(s)	IIA, 7.2.1.1.2 degradation	Supplementary soil degradation studies - anaerobic
2.	Location in Dossier	Section 5,	
3.	Authors (year) Title Report No., Date Syngenta File N° Owner	Clark, A.(1998b) Anaerobic Aquatic ABR-96098, Marc 293343/0467 Syngenta Crop Pro	
4.	Testing facility	Novartis Crop Prot Environmental Saf Greensboro, NC, U	ety Department
5.	Dates of work	-	
6.	Test substance	ISO common name  Batch:  14C-labelled test su  Specific activity of  Radiochemical pur	bstance company code: CGA 293343  Guanidinyl-4-14C label
7.	Test method	USA. Field soils w European guideline	data requirement 40 CFR 158, Subdivision N: Series 162-3, EPA, ere used as aquatic sediments. Therefore, this design meets essentially as as required in: Commission Directive 95/36/EC of 14 July 1995 Directive 91/414/EEC: Annex I: 7.1.1.1.2. Supplementary Studies; attion.
8.	Deviations	No deviations have	to be reported
9.	GLP	yes (Novartis Crop	Protection, Inc., Greensboro, NC, USA)

Test system: This second anaerobic aquatic metabolism study was performed under the identical conditions as the first study (Clark 1998a) but using guanidine-<sup>14</sup>C labelled thiamethoxam. The soil characteristics were the same as in the first study same soil as used in aerobic experiment, microbial biomass measurements only relevant for aerobic incubation).

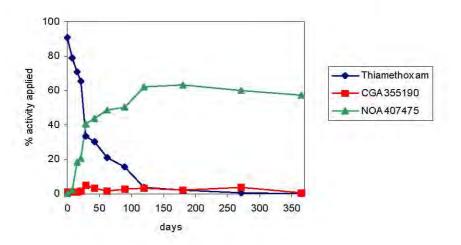
Findings: The mean total applied radioactivity recovered from the sample matrices (material balance) ranged from 98.7% to 107.3% for the kinetic samples and from 89.6% to 98.0% for the bulk samples. The distribution of radioactivity and the formation and decline of major metabolites is shown in Table 1 and Figure 1, respectively. Guanidine -<sup>14</sup>C labelled thiamethoxam was metabolised quite rapidly under anaerobic conditions to form the denitro compound NOA 407475 as a major metabolite with a peak value of 62.3% at day 120 and a decrease to 57.5% at day 365 (kinetic experiment). For the bulk experiment the peak was reached at day 62 with 58.4% and a decrease to 45.3% at day 365. The hydrolysis product CGA 355190 was also present peaking with 4.8% (kinetic) and 15.2% (bulk) at day 62 before its decline to 0.6% and 2.1%, respectively. All remaining components were below 3.8% of the dose. Some of the metabolites characterised by co-chromatography were CGA 322704, CGA 353968 and NOA 404617. Volatiles reached a maximum of 7.1% (kinetic) and 6.7% (bulk) of applied radioactivity and were identified as carbon dioxide. Non - extractables reached 22.4% (kinetic) and 12.6% (bulk).

Table 1: Distribution of radioactivity after application of <sup>14</sup>C-Guanidine-thiamethoxam in sandy loam soil, anaerobic conditions, as percent of applied dose (Clark, 1998b).

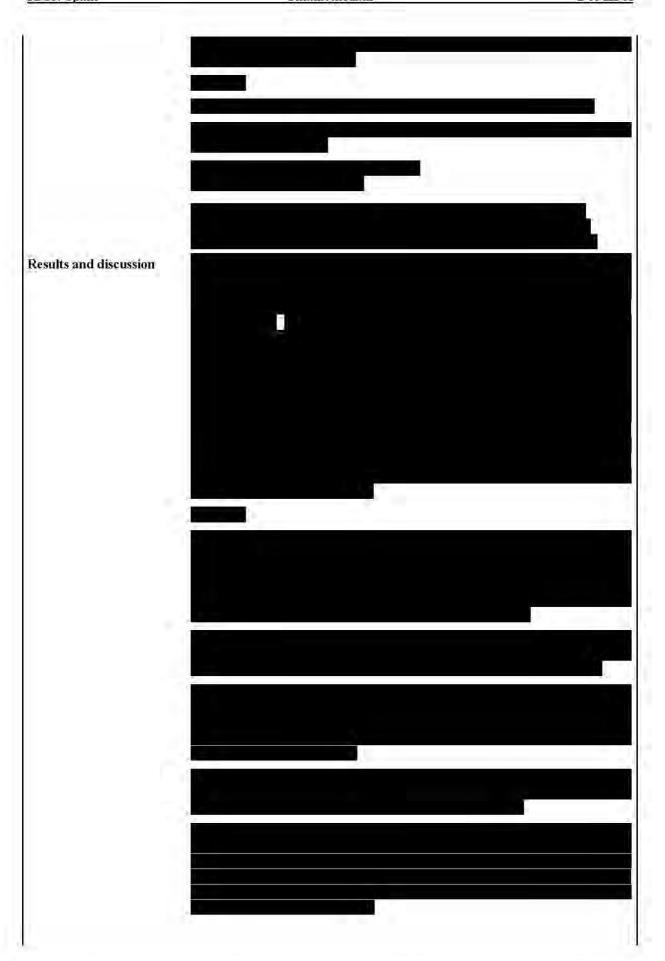
Incubation time [days]	Thiamethoxam [%]	CGA355190 [%]	NOA407475 [%]	Volatiles [%]	Non extractables [%]
0	90.9	1.0	< LOD	< LOD	0.9
7	79.1	1.0	2.1	0.5	5.0
15	70,7	1.3	18.4	1.0	5.3
21	65.6	1.7	20.5	1.8	5.2
29	33.4	4.8	40.7	1.4	5.8
43	30.3	3.0	43.6	1.3	7.7
62	21.2	1.6	48.5	2.4	18.2
90	15.9	2.6	50.2	3.6	21.8
120	4.0	3.0	62.3	7.1	14.8
180	2.0	2.3	63.4	4.5	22.1
272	0.6	3.6	59.8	5.6	19.0
365	< LOD	0.6	57.5	5.3	22.4

LOD: Level of detectability

Figure 1: Formation and decline of major metabolites for <sup>14</sup>C- Guanidine - thiamethoxam in a sandy loam soil under anaerobic conditions (Clark, 1998b)



	Evaluation by Competent Authorities	
	Use separate "evaluation boxes" to provide transparency as to the comments and views submitted	
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Date	24/09/2004	
Materials and Methods		





98/8 section	98/8 Doc IIIA 7.1.2.2.1 section No. / 01		a and	Aerobic aquatic degradation study
91/414		nnex	II	Water / sediment study
Point a	addresse	ed	7.2.1.3.2	

1.2 Title Paddy Soil Metabolism of <sup>14</sup>C- Thiazolring Labeled CGA 293343 under Laboratory

Conditions

 1.3 Report and/or project N° Syngenta File N°(Desire)
 96DA04

 293343/452

 1.4 Lab. Report N°
 96DA04

 1.5 Location in Dossier
 Section 5

1.6 AuthorsReport:Adam D.Summary:Adam D.

1.7 Date of report December 15, 1997

1.8 Published / owner Unpublished/Syngenta Crop Protection AG, Basel, Switzerland

2.1 Testing facility Novartis Crop Protection AG

**2.2 Dates of experimental work** September 26, 1996 until November 20, 1997

3. Objectives To provide information on the rate of dissipation and metabolism of <sup>14</sup>C-CGA 293343 and

its degradation products in a model paddy system under aerobic conditions.

4.1 Test substance CGA 293343

**4.2 Specification** labeled in the thiazolring

NO<sub>2</sub> N S \* C

4.3 Storage stability stable at - 20 °C

**4.4 Stability in vehicle** the test substance was found to be stable when analysed by HPLC before and after

treatmen

**4.5** Homogeneity in vehicle the test substance was prepared as a homogeneous solution in acetone

4.6 Validity not applicable

4.7 Vehicle / solvent acetone (Merck Darmstadt, FRG)

4.8 Physical form crystalline

5.1 Test method Guidedance on toxicology study data for application of agricultural chemical registration,

59 Nohsan No. 4200, January 28, 1985, Ministry of Agriculture, Forestry and Fisheries.

**5.2 Justification** The study was designed to meet Japaneser equirements to assess the metabolism of CGA

293343 in paddy soil.

5.3 Copy of method available on request

6 Choice of method not applicable

7 Deviations none 8.1 Certified laboratory yes

**8.2 Certifying authority** Federal Department of Interior, Switzerland

**8.3 GLP** GLP Switzerland based on OECD

88.4 Justification not applicable9.1 GEP not applicable

9.2 Type of facility (official or officially recognised)

not applicable

9.3 Justification

not applicable

10 Test system

The metabolism of  $^{14}$ C-thiazolring labeled CGA 293343 was investigated in a Japanese paddy soil (Ono Station, Hyogo, Japan see Table 1 for soil characteristics). The concentration of the test compound was 0.51 mg/kg dry soil, corresponding to 2.2 times the max recommended single field application rate of 0.3 kg a.i./ha, assuming a homogeneous distribution of the test substance in the top 10 cm of the soil and a soil density of 1.3 g/ml. The incubation conditions were aerobic, dark,  $25 \pm 2$  °C 363 days of incubation

Sediment characteristics:								
source	Ono Station, Hyogo Japan							
sand [%]:	47.1	total nitrogen [%]:	0.18					
silt [%]:	35.8	organic carbon [%]:	1.9					
clay [%]:	17.1	CEC [mVal/100g]:	17.8					
pН	5.0	Biomass [mg C/100g dry soil] at start	135.4					
1		at	140.8					
		end						

11 Statistics DT 50 and DT 90 values were calculated using first order one and two compartment

kinetics

12 References (published) none

13 Unpublished data No references to unpublished data were made in this summary

Over the course of the experiment an increasing proportion of radioactivity in Findings: the water phase was translocated into the sediment. Bound residues in the soil increased continuously reaching a level of 61.9% at the end of the study. Fractioning of the bound radioactivity after treatment with 0.5N NaOH at RT left 8.5% with fulvic acids and 0.6% with humic acids. The remaining radioactivity in soil could not be allocated. Carbon dioxide evolution was modest and amounted to 3.6% of applied radioactivity at day 363. The overall recovery of the system was 97.1±5.3%. Dissipation of the parent molecule from the water phase proceeded via adsorption to the soil matrix where the compound was reduced yielding the denitro metabolite NOA 407475 with a maximum of 37.1% at day 120 and a decrease to 26.9% at the end of the study. NOA 407475 was only extracted from the sediment by applying harsh extraction methods (acetonitrile/water 4/1 and acetonitrile/0.1N HCl 9/1 under reflux for two hours each). Metabolite CGA 355190 was sporadically observed in the soil at levels below the limit of quantification (0.85% of applied dose). No metabolite could be detected in the water phase at any time. The distribution of radioactivity at the various time intervals is shown in Table 1. Based on the concentrations in water and in sediment half-lives ( $DT_{50}$ ) and  $DT_{90}$  values were determined by applying first order one and two compartment reaction kinetics (Table 2).

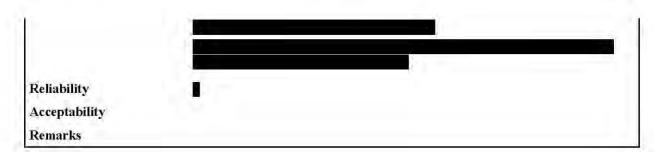
Table 1: Radioactivity distribution of thiazolyl-<sup>14</sup>C thiamethoxam and metabolites in paddy soil system as percent of applied dose

Time days	Water Layer	Sediment Extractables		Non Extractable	CO <sub>2</sub>	Recovery
,	Thiamethoxam	thiamethoxam	NOA407475			
0	97.5	5.3	<lod< td=""><td>0.6</td><td><u> </u></td><td>104.5</td></lod<>	0.6	<u> </u>	104.5
3	51.8	48.3	<lod< td=""><td><lod< td=""><td>0.06</td><td>101.1</td></lod<></td></lod<>	<lod< td=""><td>0.06</td><td>101.1</td></lod<>	0.06	101.1
8	37.7	59.6	0.6	1.1	0.27	99.6
16	27.3	69.7	1.0	1.6	0.24	99.8
42	10.5	58.8	16.9	9.4	0.42	96.0
58	6.2	36.3	29.9	20.5	2.65	95.6
120	1.6	20.0	37.1	35.1	0.99	95.6
182	0.6	7.2	28.6	45.2	3.11	85.6
363	<lod< td=""><td>2.0</td><td>26.9</td><td>61.9</td><td>3.57</td><td>96.4</td></lod<>	2.0	26.9	61.9	3.57	96.4

Table 2: Half-lives of thiazolyl-14C thiamethoxam and NOA 407475 in paddy system

	Total System		Water		Sediment	
(days)	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>
thiamethoxam	51.6	162.3	3.3	43.7	46.6	154.8
NOA 407475	-	-	-	=	330	1097

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	Evaluation by Competent Authorities
	Use separate "evaluation boxes" to provide transparency as to the comments and views submitted
	EVALUATION BY RAPPORTEUR MEMBER STATE
Date	23/8/04
Materials and Methods	Comments:
	•
Results and discussion	Comments:
Conclusion	



98/8 section	section No. / 02			Aerobic aquatic degradation study
91/414	An	inex	П	Water / sediment study
Point a	addresse	d	7.2.1.3.2	

1.2 Title Paddy Soil Metabolism of <sup>14</sup>C- Oxadiazin ring Labeled CGA 293343 under Laboratory

Conditions

1.3 Report and/or project N° 96DA05 Syngenta File N°(Desire) 293343/402

1.4 Lab. Report N° 96DA051.5 Location in Dossier Section 5

 1.6 Authors
 Report:
 Adam D.

 Summary:
 Adam D.

1.7 Date of report January 7, 1998

1.8 Published / owner Unpublished/Syngenta Crop Protection AG, Basel, Switzerland

2.1 Testing facility Novartis Crop Protection AG

**2.2 Dates of experimental** September 26, 1996 until November 20, 1997

work

3. Objectives To provide information on the rate of dissipation and metabolism of <sup>14</sup>C-CGA 293343 and

its degradation products in a model paddy system under aerobic conditions.

4.1 Test substance CGA 293343

**4.2 Specification** labeled in the oxadiazinring

4.3 Storage stability stable at - 20 °C

4.4 Stability in vehicle the test substance was found to be stable when analysed by HPLC before and after

treatment

**4.5** Homogeneity in vehicle the test substance was prepared as a homogeneous solution in acetone

4.6 Validity not applicable

4.7 Vehicle / solvent acetone (Merck Darmstadt, FRG)

4.8. Physical form crystalline

**5.1 Test method** Guidance on toxicology study data for application of agricultural chemical registration, 59

Nohsan No. 4200, January 28, 1985, Ministry of Agriculture, Forestry and Fisheries.

**5.2 Justification** The study was designed to meet Japanese requirements to assess the metabolism of CGA

293343 in paddy soil.

5.3 Copy of method available on request

6 Choice of method not applicable

7 Deviations none8.1 Certified laboratory yes

**8.2 Certifying authority** Federal Department of Interior, Switzerland

**8.3 GLP** GLP Switzerland based on OECD

8.4 Justification not applicable9.1 GEP not applicable

9.2 Type of facility (official not applicable or officially recognised)

9.3 Justification not applicable

10 Test system The metabolism of <sup>14</sup>C-oxadiazinring labeled CGA 293343 was investigated in a Japanese

paddy soil (Ono Station, Hyogo, Japan see Table 1 for soil characteristics). The concentration of the test compound was 0.52 mg/kg dry soil, corresponding to 2.2 times the max. recommended single field application rate of 0.3 kg a.i./ha, assuming a homogeneous distribution of the test substance in the top 10 cm of the soil and a soil density of 1.3 g/ml.

The incubation conditions were aerobic, dark,  $25 \pm 2$  °C 363 days of incubation.

11 Statistics DT 50 and DT 90 values were calculated using first order one and two compartment

kinetics

12 References (published) none

13 Unpublished data No references to unpublished data were made in this summary

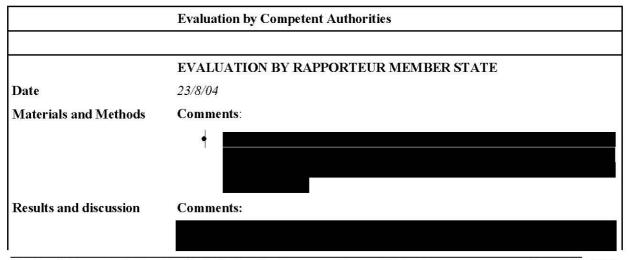
Findings: Results of this study were nearly identical to the study performed with the <sup>14</sup>C-thiazolyl labeled test substance. They are summarized in Table 1 and in Table 2.

Table 1: Radioactivity distribution of guanidine - <sup>14</sup>C thiamethoxam and metabolites in paddy soil system as percent of applied dose

Time days	Water Laver	Water Sediment Layer Extractables		Non Extractable	CO <sub>2</sub>	Recovery
uuyo		thiamethoxam	CONTRACTOR CONTRACTOR	Extraolabio	332	iteoovery
0	91.2	6.7	<lod< td=""><td>1.1</td><td>띹</td><td>100.5</td></lod<>	1.1	띹	100.5
3	52.5	47.5	1.0	0.9	0.04	101.9
8	33.8	60.0	1.8	1.8	0.09	97.9
16	26.3	65.5	4.7	3.0	0.27	99.7
42	11.6	56.4	18.3	13.3	0.87	100.4
58	6.7	40.9	30.8	20.4	1.00	99.7
120	1.8	16.6	39.1	40.3	1.21	99.0
182	0.4	10.5	39.0	43.6	1.28	94.8
363	<lod< td=""><td>1.9</td><td>30.8</td><td>62.8</td><td>2.15</td><td>99.3</td></lod<>	1.9	30.8	62.8	2.15	99.3

Table 2: Half-lives of guanidine -14C thiamethoxam and NOA 407475 in paddy system

יר או האינט ובער בארגע בער בער בער בער בער בער בער בער בער בע	Total System		Water		Sediment	
(days)	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	$DT_{90}$
thiamethoxam	51.8	170	3.4	47.1	39.2	130.2
NOA 407475	7 <b>2</b> 1	=	2	324	159	529





98/8 Doc IIIA 7.1.2.2.2 section No. / 01		Water sediment degradation study
91/414 Annex Point addressed	II 7.2.1.3.2	Water / sediment study

Title 1.2 Degradation and Metabolism of <sup>14</sup>C-oxadiazinring Labeled CGA 293343 in two Aerobic

Aquatic Systems under Laboratory Conditions

Report and/or project N° 96DA02 Syngenta File N°(Desire) 293343/436

1.4 Lab. Report Nº 96DA02. Location in Dossier Section 5

16 Authors Report: Adam D. Adam D. Summary:

1.7 Date of report February 4, 1998

1.8 Published / owner Unpublished/Syngenta Crop Protection AG, Basel, Switzerland

2.1 **Testing facility** Novartis Crop Protection AG

2.2 Dates of experimental

work

August 20, 1996 until November 13, 1997

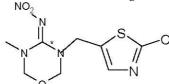
3. **Objectives** To provide information on the rate of dissipation and metabolism of  $^{14}\mathrm{C\text{-}CGA}$  293343 and

its degradation products in two equilibrated water/sediment systems under aerobic

conditions.

CGA 293343 Test substance 4.1

4.2 **Specification** labeled in the oxadiazinring



4.3 Storage stability stable at - 20 °C

4.4 Stability in vehicle the test substance was found to be stable when analysed by HPLC before and after treatment

4.5 Homogeneity in vehicle the test substance was prepared as a homogeneous solution in acetonitrile

4.6 Validity not applicable

Vehicle / solvent acetonitrile (Merck Darmstadt, FRG)

4.8 Physical form crystalline

Commission Directive 95/36/EC of July 14, 1995 amending Council Directive 91/414/EEC: 5.1 Test method

Annex II: 7.2.1.3.2. Water/Sediment Study; referencing

Procedures for Assessing the Environmental Fate and Ecotoxicity of Pesticides, 8.2. Aerobic Aquatic Degradation, Society of Environmental Toxicology and Chemistry, SETAC Europe Richtlinie für die Prüfung von Pflanzenschutzmitteln, Teil IV, 5-1, Abbaubarkeit und Verbleib von Pflanzenschutzmitteln im Wasser/Sediment-System. Biologische

Bundesanstalt für Land- und Forstwirtschaft Bundesrepublik Deutschland, Dezember 1990.

5.2 Justification The study was designed to meet several requirements to assess the metabolism of CGA

293343 in water/sediment systems.

5.3 Copy of method available on request Choice of method not applicable

**Deviations** 7 none 8.1 Certified laboratory ves

8.2 Certifying authority Federal Department of Interior, Switzerland **8.3 GLP** GLP Switzerland based on OECD

8.4 Justification not applicable
 9.1 GEP not applicable
 9.2 Type of facility (official or officially recognised)

**9.3 Justification** not applicable

10 Test system

The metabolism of <sup>14</sup>C-oxadiazinring labeled CGA 293343 was investigated in two

equilibrated water sediment systems (see Tables 1 and 2 for water and sediment characteristics). The concentration of the test compound was 0.1 mg/kg with respect to the water phase, corresponding to 1.5 times the max. recommended single field application rate of 0.2 kg a.i./ha (assuming a homogeneous distribution of the test substance to a water depth of 30 cm). The incubation conditions were aerobic, dark,  $20 \pm 2$  °C, 100 days of incubation.

11 Statistics DT 50 and DT 90 values were calculated using first order one and two compartment kinetics

12 References (published) none

13 Unpublished data Adam, D., Paddy Soil Metabolism of <sup>14</sup>C-oxadiazinring Labeled CGA 293343 under

Laboratory Conditions. Project report 96DA05 (1998) Syngenta Crop Protection AG,

Environmental Safety / Ecochemistry, 4002 Basel, Switzerland

Findings: Results of the radioactivity distribution between water and sediment and formation and decline of the major metabolite NOA 407475, bound residues and carbon dioxide were very similar to the study performed with the thiazolyl- $^{14}$ C labelled test compound. The distribution and nature of radioactivity for the incubation period in the river and pond water-sediment systems are shown in Table 1 and 2. The denitro metabolite NOA 407475 reached maximum concentrations of 40.9% at day 58 and of 47.4% at day 100 in the river and pond sediment, respectively. The metabolite CGA 355190 was observed in both water and sediment of the 2 systems at levels <5% of applied dose. Half-lives and  $DT_{90}$  values for thiamethoxam are presented in Table 7.2.1.3.2 - 7.

Table 1: Radioactivity distribution of guanidine-<sup>14</sup>C thiamethoxam and metabolites in river aquatic system as percent of applied dose

Time days		ater yer		Sediment Extractables		Non Extractable	CO <sub>2</sub>	Recover y
	thiamethoxa	CGA 355190	thiamethoxa	CGA 355190	NOA40747			
0	88.7	<lod< td=""><td>12.4</td><td><lod< td=""><td><lod< td=""><td>0.9</td><td><b>E</b>(1)</td><td>102.1</td></lod<></td></lod<></td></lod<>	12.4	<lod< td=""><td><lod< td=""><td>0.9</td><td><b>E</b>(1)</td><td>102.1</td></lod<></td></lod<>	<lod< td=""><td>0.9</td><td><b>E</b>(1)</td><td>102.1</td></lod<>	0.9	<b>E</b> (1)	102.1
1	55.0	<lod.< td=""><td>25.6</td><td><lod< td=""><td>2.3</td><td>10.0</td><td>0.03</td><td>92.9</td></lod<></td></lod.<>	25.6	<lod< td=""><td>2.3</td><td>10.0</td><td>0.03</td><td>92.9</td></lod<>	2.3	10.0	0.03	92.9
3	57.3	<lod< td=""><td>32.5</td><td><lod< td=""><td><lod< td=""><td>4.6</td><td>0.04</td><td>94.4</td></lod<></td></lod<></td></lod<>	32.5	<lod< td=""><td><lod< td=""><td>4.6</td><td>0.04</td><td>94.4</td></lod<></td></lod<>	<lod< td=""><td>4.6</td><td>0.04</td><td>94.4</td></lod<>	4.6	0.04	94.4
8	44.2	<lod.< td=""><td>36.6</td><td>1.1</td><td>8.6</td><td>2.6</td><td>0.06</td><td>94.9</td></lod.<>	36.6	1.1	8.6	2.6	0.06	94.9
16	39.7	<lod< td=""><td>32.1</td><td>2.5</td><td>15.9</td><td>4.7</td><td>0.88</td><td>95.8</td></lod<>	32.1	2.5	15.9	4.7	0.88	95.8
42	24.7	<lod< td=""><td>24.1</td><td>2.9</td><td>34.5</td><td>11.5</td><td>1.42</td><td>102.4</td></lod<>	24.1	2.9	34.5	11.5	1.42	102.4
58	16.2	1.0	13.3	2.6	40.9	15.5	3.03	92.5
80	11.4	2.5	12.6	4.0	34.6	22.2	4.89	92.3
100	13.6	4.5	11.5	4.4	35.9	19.8	7.22	97.8

Table 2: Radioactivity distribution of guanidine-<sup>14</sup>C thiamethoxam and metabolites in pond aquatic system as percent of applied dose

Time days	Water Layer		Sediment Extractables			Non Extractable	CO <sub>2</sub>	Recover
	thiamethoxa	CGA 355190	thiamethoxa	CGA 355190	NOA407475			
0	97.8	<lod< td=""><td>3.7</td><td><lod< td=""><td><lod< td=""><td>0.6</td><td></td><td>102.1</td></lod<></td></lod<></td></lod<>	3.7	<lod< td=""><td><lod< td=""><td>0.6</td><td></td><td>102.1</td></lod<></td></lod<>	<lod< td=""><td>0.6</td><td></td><td>102.1</td></lod<>	0.6		102.1
1	80.6	<lod< td=""><td>13.4</td><td><lod< td=""><td><lod< td=""><td>2.8</td><td>0.00</td><td>96.7</td></lod<></td></lod<></td></lod<>	13.4	<lod< td=""><td><lod< td=""><td>2.8</td><td>0.00</td><td>96.7</td></lod<></td></lod<>	<lod< td=""><td>2.8</td><td>0.00</td><td>96.7</td></lod<>	2.8	0.00	96.7
3	74.4	<lod< td=""><td>20.1</td><td><lod< td=""><td><lod< td=""><td>2.5</td><td>0.03</td><td>97.0</td></lod<></td></lod<></td></lod<>	20.1	<lod< td=""><td><lod< td=""><td>2.5</td><td>0.03</td><td>97.0</td></lod<></td></lod<>	<lod< td=""><td>2.5</td><td>0.03</td><td>97.0</td></lod<>	2.5	0.03	97.0
9	54.6	2.3	24.2	1.3	2.3	7.8	0.00	93.7

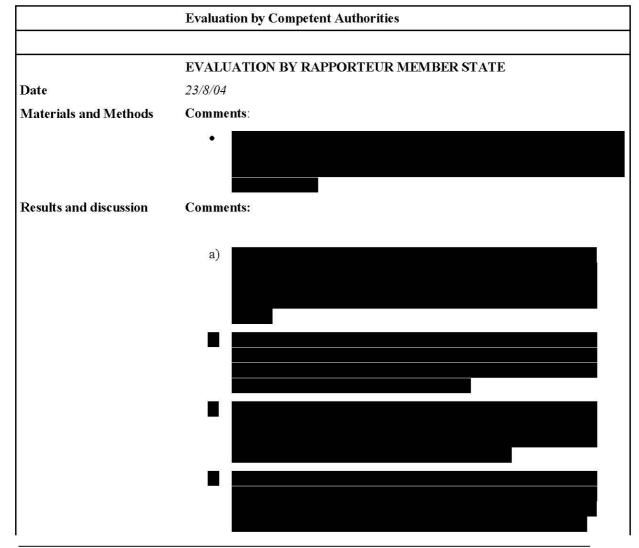
16	*	*	*	*	*	*	*	*
42	21.3	3.6	15.7	3.7	31.6	10.4	2.74	88.9
58	10.6	3.0	15.2	2.7	34.8	14.4	3.12	85.4
80	7.5	1.4	10.6	3.3	40.2	20.2	7.18	90.4
100	6.4	<lod< td=""><td>10.4</td><td>3.2</td><td>47.4</td><td>23.9</td><td>5.97</td><td>98.2</td></lod<>	10.4	3.2	47.4	23.9	5.97	98.2

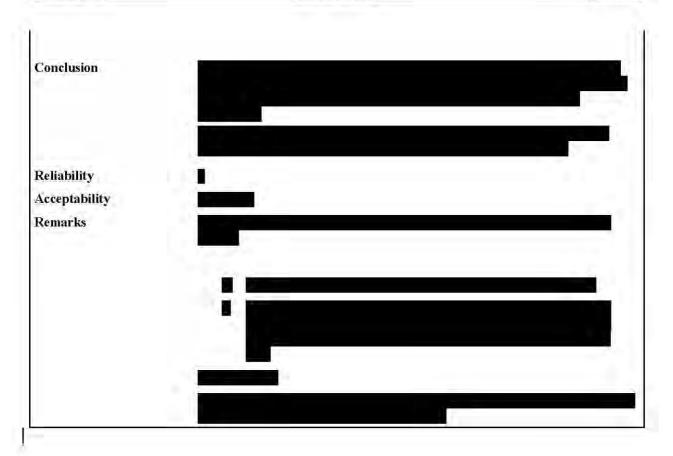
<sup>\*</sup> Data for day 16 disregarded due to double application

Table 3: Half-lives of guanidine -14C thiamethoxam in aquatic systems

	Total System			ater	Sediment	
(days)	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>
River	35.3	615.6	5.8	89.2	11.5	>100
Pond	28.4	325.4	13.2	69.7	9.8	>100

As discussed for the aquatic study above (Adam,1998a), distribution coefficients  $k_d$  for any time can be estimated based on the relative distribution of thiamethoxam to water/sediment in Table 1 and Table 2 for Rhine river and pond aquatic systems, respectively. 55  $\mu g^{-14}C$ -thiamethoxam were applied per flask. After 42 days of incubation the concentrations of  $^{14}C$ -thiamethoxam in the sediments can be calculated to be about 65  $\mu g/kg$  and 44  $\mu g/kg$  in the river and pond and the concentration in water can be calculated to about 25  $\mu g/litre$  and 21  $\mu g/litre$ , respectively. Thus, the mean sediment/water distribution coefficient can be estimated to be  $k_d = 2.6$  and 2.1 mL/g.





98/8 section	Doc n No.	IIIA	7.1.2.2.2 <i>I</i> 02	Water sediment degradation study
91/414		Annex	П	Water / sediment study
Point a	addres	sed	7.2.1.3.2	100

1.2 Title Degradation and Metabolism of 14C-thiazolring Labeled CGA 293343 in two Aerobic

Aquatic Systems under Laboratory Conditions

1.3 Report and/or project N° 96DA01 Syngenta File N°(Desire) 293343/401

1.4 Lab. Report N° 96DA01
1.5 Location in Dossier Section 5

 1.6 Authors
 Report:
 Adam D.

 Summary:
 Adam D.

1.7 Date of report January 9, 1998

1.8 Published / owner Unpublished/Syngenta Crop Protection AG, Basel, Switzerland

2.1 Testing facility Novartis Crop Protection AG

2.2 Dates of experimental Jun

work

June 4, 1996 until November 13, 1997

3. Objectives To provide information on the rate of dissipation and metabolism of <sup>14</sup>C-CGA 293343 and

its degradation products in two equilibrated water/sediment systems under aerobic

conditions.

4.1 Test substance CGA 293343

**4.2 Specification** labeled in the thiazolring

NO<sub>2</sub>N S \* C

4.3 Storage stability stable at - 20 °C

**4.4 Stability in vehicle** the test substance was found to be stable when analysed by HPLC before and after

treatment

**4.5** Homogeneity in vehicle the test substance was prepared as a homogeneous solution in acetonitrile

**4.6 Validity** not applicable

4.7 Vehicle / solvent acetonitrile (Merck Darmstadt, FRG)

4.8 Physical form crystalline

5.1 Test method Commission Directive 95/36/EC of July 14, 1995 amending Council Directive

91/414/EEC: Annex II: 7.2.1.3.2. Water/Sediment Study; referencing

Procedures for Assessing the Environmental Fate and Ecotoxicity of Pesticides, 8.2. Aerobic Aquatic Degradation, Society of Environmental Toxicology and Chemistry,

SETAC Europe

Richtlinie für die Prüfung von Pflanzenschutzmitteln, Teil IV, 5-1, Abbaubarkeit und Verbleib von Pflanzenschutzmitteln im Wasser/Sediment-System. Biologische

Bundesanstalt für Land- und Forstwirtschaft Bundesrepublik Deutschland, Dezember 1990.

**5.2 Justification** The study was designed to meet several requirements to assess the metabolism of CGA

293343 in water/sediment systems.

5.3 Copy of method available on request6 Choice of method not applicable

7 Deviations none

Certified laboratory yes

8.2 Certifying authority Federal Department of Interior, Switzerland

8.3 GLP GLP Switzerland based on OECD

Justification 8.4 not applicable 9.1 **GEP** not applicable 9.2 Type of facility (official not applicable or officially recognised)

9.3 Justification

**Test system** 

10

not applicable

The metabolism of <sup>14</sup>C-thiazolring labeled CGA 293343 was investigated in two equilibrated water sediment systems (see Tables 1 and 2 for water and sediment characteristics). The concentration of the test compound was 0.1 mg/kg with respect to the water phase, corresponding to 1.5 times the max. recommended single field application rate of 0.2 kg a.i./ha (assuming a homogeneous distribution of the test substance to a water depth to 30 cm). The incubation conditions were aerobic, dark,  $20 \pm 2$  °C, 100 days of

incubation.

Water/sediment characteristics of river and pond systems

Sediment characteristics:	River (Rhine)	Pond (Rheinfelden)
sand [%]:	41.8	30.9
silt [%]:	48.4	41.3
clay [%]:	9.8	27.8
рН	7.5	-
total nitrogen [%]:	0.1	0.19
organic carbon [%]:	1.7	2.2
CEC [mVal/100g]:	10.9	17.8
Biomass [mg C/100g dry soil]:	78	119
Water characteristics	an and the second of the secon	
рН	8.25	7.9
Oxygen content (mg/l)	0.1	0.9
TOC (mg/l)	2.5	5.6
total nitrogen (mg/l)	3.3	3.6

**Statistics** 11 DT 50 and DT 90 values were calculated using first order one and two compartment

kinetics

none References (published) 12

13 Unpublished data Adam, D., Paddy Soil Metabolism of <sup>14</sup>C-Thiazolring Labeled CGA 293343 under

Laboratory Conditions. Project report 96DA04 (1997) Syngenta Crop Protection AG,

Environmental Safety / Ecochemistry, 4002 Basel, Switzerland

Findings: In both systems the radioactivity in the water phase decreased over the incubation period reaching 18% (river) and 6% (pond) of the applied radioactivity at day 100. In the sediment the amount of non extractable radioactivity increased continuously to a level of 13.8% (river) and 15.0% (pond). All volatile radioactivity was characterized as carbon dioxide. It reached 6.3% (river) and 9.3% (pond) at day 100, indicating some mineralisation. The distribution and nature of radioactivity for the incubation period in the river and pond watersediment systems are shown in Table 1 and 2. Dissipation of the parent molecule from the water phase proceeded mainly via adsorption to the sediment matrix where the compound was reduced to the denitro metabolite NOA 407475. NOA 407475 could be extracted by a harsh extraction method using acetonitrile-water and acetonitrile-hydrochloric acid under reflux. It reached maximum concentrations of 37.0% at day 58 and decreased to 35.4% at day 100 in the river and of 47.4% at day 42 and of 45.5% at day 100 in the pond sediment, respectively. The other metabolite identified as CGA 355190 was observed in both water and sediment of the river system at levels below 5% beginning at day 42. In the pond system it was present beginning at day 8 at lower levels. Based on the concentrations of the parent molecule in water and sediment, half-lives ( $DT_{50}$ ) and  $DT_{90}$  values were determined by applying first order one and two compartment reaction kinetics (Table 3).

Table 1: Radioactivity distribution of thiazolyl-<sup>14</sup>C thiamethoxam and metabolites in river aquatic system as percent of applied dose

Time days	Water Layer		Annual Control of the			Non Extractable	CO <sub>2</sub>	Recover y
	thiamethoxa	CGA 355190	thiamethoxa	CGA 355190	NOA40747			
0	102.0	<lod< td=""><td>1.5</td><td><lod< td=""><td><lod< td=""><td>0.3</td><td>engangan ang mililikan ang milipags</td><td>103.8</td></lod<></td></lod<></td></lod<>	1.5	<lod< td=""><td><lod< td=""><td>0.3</td><td>engangan ang mililikan ang milipags</td><td>103.8</td></lod<></td></lod<>	<lod< td=""><td>0.3</td><td>engangan ang mililikan ang milipags</td><td>103.8</td></lod<>	0.3	engangan ang mililikan ang milipags	103.8
1	81.5	<lod.< td=""><td>13.5</td><td><lod< td=""><td><lod.< td=""><td>0.4</td><td>0.02</td><td>95.3</td></lod.<></td></lod<></td></lod.<>	13.5	<lod< td=""><td><lod.< td=""><td>0.4</td><td>0.02</td><td>95.3</td></lod.<></td></lod<>	<lod.< td=""><td>0.4</td><td>0.02</td><td>95.3</td></lod.<>	0.4	0.02	95.3
3	78.1	<lod< td=""><td>23.4</td><td><lod< td=""><td><lod< td=""><td>2.0</td><td>0.09</td><td>103.6</td></lod<></td></lod<></td></lod<>	23.4	<lod< td=""><td><lod< td=""><td>2.0</td><td>0.09</td><td>103.6</td></lod<></td></lod<>	<lod< td=""><td>2.0</td><td>0.09</td><td>103.6</td></lod<>	2.0	0.09	103.6
8	63.6	<lod.< td=""><td>27.2</td><td><lod.< td=""><td><lod< td=""><td>5.3</td><td>0.88</td><td>98.2</td></lod<></td></lod.<></td></lod.<>	27.2	<lod.< td=""><td><lod< td=""><td>5.3</td><td>0.88</td><td>98.2</td></lod<></td></lod.<>	<lod< td=""><td>5.3</td><td>0.88</td><td>98.2</td></lod<>	5.3	0.88	98.2
16	37.2	<lod< td=""><td>33.6</td><td><lod< td=""><td>18.8</td><td>5.1</td><td>1.40</td><td>96.0</td></lod<></td></lod<>	33.6	<lod< td=""><td>18.8</td><td>5.1</td><td>1.40</td><td>96.0</td></lod<>	18.8	5.1	1.40	96.0
42	21.9	3.0	21.9	<lod< td=""><td>33.0</td><td>7.5</td><td>1.07</td><td>94.1</td></lod<>	33.0	7.5	1.07	94.1
58	22.9	1.5	13.6	4.1	37.0	5.8	1.56	94.0
80	16.2	3.5	12.7	4.0	30. 7	12.6	2.41	88.0
100	11.8	4.0	12.2	4.7	35.4	13.8	6.34	90.9

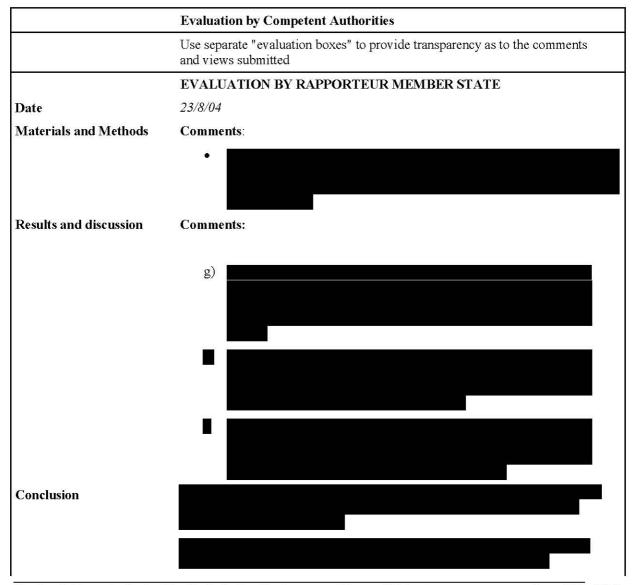
Table 2: Radioactivity distribution of thiazolyl-<sup>14</sup>C thiamethoxam and metabolites in pond aquatic system as percent of applied dose

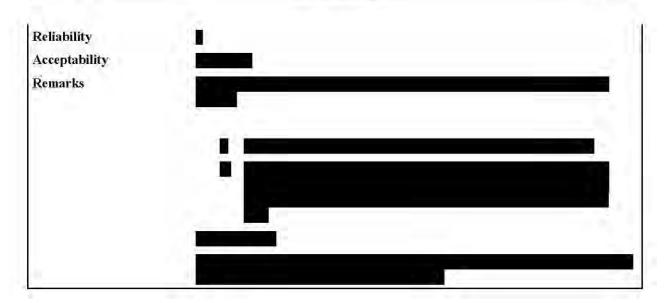
Time days	Water Layer			Sediment Extractables			CO <sub>2</sub>	Recovery
	thiamethoxa	CGA 355190	thiamethoxa	CGA 355190	NOA407475			
0	104.7	<lod< td=""><td>2.4</td><td><lod< td=""><td><lod< td=""><td>0.7</td><td></td><td>107.8</td></lod<></td></lod<></td></lod<>	2.4	<lod< td=""><td><lod< td=""><td>0.7</td><td></td><td>107.8</td></lod<></td></lod<>	<lod< td=""><td>0.7</td><td></td><td>107.8</td></lod<>	0.7		107.8
1	85.2	<lod< td=""><td>14.9</td><td><lod< td=""><td><lod< td=""><td>1.0</td><td>0.02</td><td>101.1</td></lod<></td></lod<></td></lod<>	14.9	<lod< td=""><td><lod< td=""><td>1.0</td><td>0.02</td><td>101.1</td></lod<></td></lod<>	<lod< td=""><td>1.0</td><td>0.02</td><td>101.1</td></lod<>	1.0	0.02	101.1
3	75.9	<lod< td=""><td>21.2</td><td><lod< td=""><td><lod< td=""><td>3.5</td><td>0.07</td><td>100.6</td></lod<></td></lod<></td></lod<>	21.2	<lod< td=""><td><lod< td=""><td>3.5</td><td>0.07</td><td>100.6</td></lod<></td></lod<>	<lod< td=""><td>3.5</td><td>0.07</td><td>100.6</td></lod<>	3.5	0.07	100.6
8	53.7	2.2	22.3	1.2	15.1	4.4	0.05	99.9
16	33.6	1.1	31.2	<lod< td=""><td>16.2</td><td>7.5</td><td>0.44</td><td>92.2</td></lod<>	16.2	7.5	0.44	92.2
42	12.3	2.1	14.5	3.4	47.4	7.1	2.19	89.0
58	26.7	1.3	13.9	2.5	32.1	21.9	5.70	106.0
80	2.97	<lod< td=""><td>9.8</td><td>2.6</td><td>46.3</td><td>25.3</td><td>3.17</td><td>88.7</td></lod<>	9.8	2.6	46.3	25.3	3.17	88.7
100	6.2	<lod< td=""><td>9.8</td><td>3.4</td><td>45. 5</td><td>15.0</td><td>9.28</td><td>89.0</td></lod<>	9.8	3.4	45. 5	15.0	9.28	89.0

Table 3: Half-lives of thiazolyl-14C thiamethoxam in aquatic systems

	Total System		Water		Sediment	
(days)	DT <sub>50</sub>	$DT_{90}$	DT <sub>50</sub>	DT <sub>90</sub>	DT <sub>50</sub>	DT <sub>90</sub>
River	36.3	317.4	11.0	132.3	12.6	41.9
Pond	25.7	130.9	8.4	76.9	15.6	51.9

Based on the relative distribution of thiamethoxam to water/sediment in Table 1 and Table 2 for Rhine river and pond aquatic systems, respectively, distribution coefficients  $k_d$  for any time can be estimated taking the dimensions of the systems into account. After decantation/centrifugation about 540 mL water phase represented the water layer. The mean dry weight of the sediments was 122 g per flask for the river and for the pond system corresponding to about 200 g wet sediments taken for extraction after decanting. 52  $\mu$ g <sup>14</sup>C-thiamethoxam were applied per flask. After 16 days of incubation equilibrium was reached in both systems and about 34% of the applied radioactivity were found in both systems in the water phase as well as in the sediment. The concentrations of <sup>14</sup>C-thiamethoxam in the sediments can be calculated to be about 89  $\mu$ g/kg and the concentration of <sup>14</sup>C-thiamethoxam in water can be calculated to about 33  $\mu$ g/litre. Thus, a mean sediment/water distribution coefficient can be estimated to be  $k_d = 2.7$  mL/g.





98/8 D section N	oc IIIA o.	7.2.2.1 / 01	The rate and route of degradation including identification of the process involved and identification of any metabolites and degradation products in at least three soil types under appropriate conditions
91/414	Annex	10	Rate of degradation in soil - laboratory studies
Point add	ressed	7.1.1.2.1 / 03	

1.2 Title RATE OF DEGRADATION OF CGA 293343 IN SOIL UNDER VARIOUS

CONDITIONS

Report and/or project N° 95RP03 Syngenta File N°(Desire)

293343/98

1.4 Lab. Report Nº 95RP03

Cross reference to original study / report

Authors Dr. R. Phaff Report: Dr. R. Phaff Summary:

Date of report 1.7 May 23, 1997a

Published / owner Owned by Syngenta Crop Protection AG, not published

2.1 **Testing facility** Novartis Crop Protection AG

Environmental Safety/Ecochemistry

4002 Basel, Switzerland

2.2 Dates of experimental

work

May 15, 1995 until January 7, 1997

The objectives of the study were to determine the rates and routes of degradation of CGA 3. Objectives

293 343 in soil incubated under various laboratory conditions...

Test substance ISO common name: thiamethoxam 4.1

Trade name:

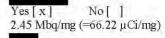
Batch:

<sup>14</sup>C-labelled test substance Specific activity of [.....]

Radiochemical purity of the test substance:

Structural formula: (position of label)

Yes [



0

No [x]

Formulation used for study: Type of formulation (if used):

Co-solvent for application (if used): acetone

4.2 Specification See 4.1

4.3 Storage stability The chemical is stable when stored at about -20°C in the dark

Stability in vehicle The test substance was found to be stable in the vehicle when analysed by TLC before and 4.4

after application

Homogeneity in vehicle The test substance was prepared as a homogeneous solution in acetonitrile

4.6 Validity Not applicable

4.5 Vehicle / solvent Acetone **4.7 Physical form** crystalline, slightly beige

**5.1 Test method** Danish Law of the Ministry of the Environment, September 1, 1987.

Dutch Registration Guideline, Section G.1: Behaviour in Soil; Ministry of Agriculture and Fisheries, Ministry of Public Health and Environmental Hygiene, Ministry of Social

Affairs, January 1987.

Environmental Chemistry and Fate Guidelines for Registration of Pesticides in Canada; Section C: Biotransformation: 1: Soil-Degradation Pathways and Persistence, July 15, 1987

5.2 Justification not applicable

5.3 Copy of method available on request6 Choice of method not applicable

7 Deviations none

8.1 Certified laboratory yes

**8.2 Certifying authority** Federal Department of Interior, Switzerland

**8.3 GLP** GLP Switzerland based on OECD

8.4 Justification not applicable
 9.1 GEP Not applicable
 9.2 Type of facility (official or officially recognised)

9.3 Justification Not applicable

## Test system

Origin (textural class)	Gartenacker (silty loam soil)
Batch Nr.	4/95
Collecting Date:	April 18, 1995
Analysis Date:	June 12, 1995
pH (KCl)	7.15
Organic Carbon	2.50 %
Total Nitrogen	0.32 %
CEC <sup>26</sup>	14.90 mmol/z/100 g soil
Particle Size Distribution -Clay (<0.002 mm) -Silt (0.002 - 0.05 mm) -Loam (0.05 - 2 mm)	11.10 % 55.00 % 33.90 %
Maximum Water Holding Capacity	69.87 g water / 100g dry soil
Field Capacity (1.8 pF)	49.17 g water / 100g dry soil
Microbial Biomass - Start - after 191 days	79 mg C/100 g dry soil) 61 mg C/100 g dry soil) at 20°C 55 mg C/100 g dry soil) at 10°C

Test conditions		Α	В	С	D
Incubation temperature(s)	(C)	20±2	20±2	10±2	20±2
Humidity	(%FC)	60	40	60	60
Treatment rate:	(mg a.i./ kg soil)	0.9	0.9	0.9	0.1
Incubation time:	(days)	363	363	363	363
Number of samples taken for analysis:		20	20	20	20
Methods used for analysis	HPLC	12	: <u></u>	-	-
Colored Colore	TLC	Х	х	Х	Х
	GC	<i>9</i>	52	<del>_</del>	ST.
Methods for the identific degradates		cochromatog	raphy, LC-MS		

**Statistics** 11 The rate of disappearance of CGA 293343 was calculated by applying first order reaction kinetics

References (published) 12 none

Unpublished data none

Findings: The metabolite formed with the highest concentration was the compound CGA 322704, N-(2-Chloro-thiazol-5-ylmethyl)-N'-methyl-N''-nitro-guanidine. Its structure was confirmed by mass spectral analysis. It reached its maximum concentration of 24%, 17% and 36 % for the experiments A, B and D on day 128, 189 and 90, respectively. Thereafter the

<sup>26</sup> Cation Exchange Capacity

concentration of the metabolite declined thus demonstrating its transient nature. In experiment C at a temperature of 10°C CGA 322704 reached a maximum of 29% at the end of the study. Details of the degradation patterns observed for incubation part A-D are summarised in Table 1.

Besides CGA 322704 only minor amounts of extractable degradation products were observed. The sum of unknown metabolites in the experiments A, B, C and D reached maximum levels of 8.1% at day 189, of 4.4% at day 128, of 3.7% at day 189 and of 8.5% at day 128, respectively with a subsequent decline. Identification of the 8.1% fractions (experiment A, day 189) by cochromatography with reference compounds showed 3 components: CGA 355190 (4.9%), CGA 265307 (2.7%) and CGA 353968 (0.7%). The data indicate that the pattern of decline is the same under all conditions but that the rate of degradation of the as and the main metabolite CGA 322704 show some dependence on incubation temperature, as concentration and soil humidity.

The mineralisation of thiazolyl-<sup>14</sup>C labelled thiamethoxam to carbon dioxide increased steadily with time in all experiments and reached a maximum of 44.2% at the low 0.11 mg/kg treatment rate. Bound residues also increased with time reaching a maximum of 19.5% after 363 days in study part B.

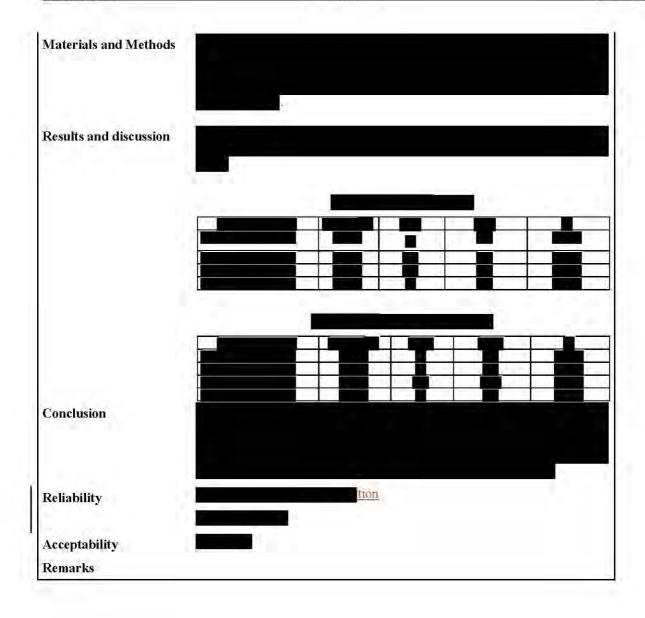
Table 1: Recovery and degradation pattern of soil degradation of <sup>14</sup>C-thiamethoxam under various conditions (Phaff 1997a)

Days after applica- tion	Volatiles (CO <sub>2</sub> ) [%]*	Extrac- tables [%]*	Extrac- tables (Soxhlet) [%]*	CGA- 293343 [%]*	CGA- 322704 [%]*	Un- known [%]*	Non- extrac- tables [%]*	Total [%]*
Part A: 20	0℃, 60% FC	C, 0.91 mg						
0	n.d.	101.1	5.6	106.7	n.d.	n.d.	0.5	107.1
3	0.1	94.4	3.7	98.1	n.d.	n.d.	1.5	99.8
7	0.5	92.2	5.7	93.5	4.4	n.d.	1.4	99.8
14	1.5	89.1	5.6	89.0	5.7	n.d.	2.2	98.4
28	3.5	83.9	5.3	81.6	7.6	n.d.	3.4	96.0
58	7.0	75.1	5.3	58.9	18.4	3.1	5.5	93.0
90	9.9	64.2	5.4	45.3	20.3	4.1	7.6	87.1
128	15.2	53.5	6.4	31.1	23.5	5.4	9.8	84.9
189	21.7	40.6	5.4	15.5	22.4	8.1	12.6	80.2
363	32.1	20.8	4.0	4.2	17.0	3.6	16.2	73.0
Part B: 20	0℃, 40% FC	C, 0.91 mg	J/kg					
0	n.d.	100.6	4.9	105.6	n.d.	n.d.	0.4	106.0
3	n.d.	93.8	3.4	97.2	n.d.	n.d.	1.5	98.7
7	0.2	92.6	6.1	94.5	4.2	n.d.	1.4	100.3
14	0.6	91.0	6.0	95.4	1.5	n.d.	1.9	99.4
28	1.0	89.1	5.2	88.9	5.3	n.d.	2.8	98.0
58	2.7	82.1	6.3	77.2	9.2	2.0	4.2	95.2
90	5.4	75.2	6.0	65.4	11.6	4.1	6.1	92.5
128	8.9	65.6	6.0	53.7	13.4	4.4	8.1	88.5
189	13.1	56.1	5.8	40.9	16.7	4.3	11.0	86.0
363	23.5	27.3	4.3	17.6	10.9	3.0	19.5	74.6
Part C: 10	0℃, 60% F0	C, 0.91 mg	j/kg				useon puseon buseon puseon buseon u	abethiasebhasebhasebhasebh
0	n.d.	101.4	2.7	106.0	n.d.	n.d.	n.d.	106.5
3	0.1	95.9	3.3	99.2	n.d.	n.d.	1.4	100.7
7	0.4	93.6	4.6	95.6	2.6	n.d.	1.2	99.7
14	0.8	92.6	5.0	96.4	1.2	n.d.	1.3	99.6
28	1.6	92.5	4.3	93.0	3.8	n.d.	1.8	100.2
58	3.2	89.2	4.5	83.0	10.8	n.d.	2.5	99.4
90	4.5	84.4	5.6	76.5	11.1	2.4	3.2	97.7
128	6.6	81.2	5.9	71.4	14.0	1.8	3.7	97.4
189	9.4	73.9	5.8	57.7	18.3	3.7	5.1	94.2
363	16.7	60.8	5.7	34.0	29.3	2.2	8.4	91.6
Part D: 20	0℃, 60% FC	C, 0.11 mg	j/kg					
0	n.d.	103.8	6.5	110.3	n.d.	n.d.	0.5	110.9
3	0.5	93.4	4.3	97.7	n.d.	n.d.	1.3	99.5
7	1.4	87.9	7.7	89.5	6.1	n.d.	1.8	98.8
14	3.5	85.6	5.4	80.8	9.4	n.d.	2.9	97.3
28	7.0	82.2	5.0	66.4	18.3	2.5	4.5	98.6
58	14.0	63.1	5.8	31.8	33.5	3.6	7.5	90.4
90	21.3	50.8	5.9	15.5	35.6	5.6	10.9	88.9
128	28.9	40.0	6.8	6.2	32.1	8.5	12.7	88.3
189	35.2	26.1	6.2	3.6	21.6	7.1	15.6	83.0
363	44.2	13.9	4.3	1.0	14.4	2.7	16.8	79.1
* Doroor			tivity avo			۷.۱	10.0 pd -	- not datas

<sup>\*</sup> Percent of applied radioactivity, average of duplicates.

n.d. = not detected

	Evaluation by Competent Authorities	
	EVALUATION BY RAPPORTEUR MEMBER STATE	
Date	26/08/04	



98/8 Doc section No.	IIIA	7.2.2.1 <i>J</i> 02	The rate and route of degradation including identification of the processes involved and identification of any metabolites and degradation products in at least three soil types under appropriate conditions
	nnex	11	Soil route of degradation: Aerobic degradation
Point address	sea	7.1.1.1.1	

1.2 Title DEGRADATION OF <sup>14</sup>C-THIAZOLRING LABELLED CGA 293343 IN VARIOUS

SOILS UNDER LABORATORY CONDITIONS

Report and/or project N° 95DA03 1.3 Syngenta File N°(Desire)

293343/141 Lab. Report Nº 1.4 95DA03 Location in dossier Section 5

Authors 1.6 Adam, D. Report: Summary: Adam, D.

1.7 Date of report December 17, 1996

Published I owner Unpublished/Syngenta Crop Protection AG, Basel, Switzerland 1.8

**Testing facility** Novartis Crop Protection AG 2.1 Product Safety / Ecochemistry 4002 Basel, Switzerland

2.2 Dates of experimental work

1.5

November 16, 1995 until November 21, 1996

3. Objectives The objectives of the study were to determine the rate of degradation in the different soils

examined.

4.1 Test substance ISO common name: thiamethoxam

Trade name:

Batch:

<sup>14</sup>C-labelled test substance Specific activity of [.....]

Radiochemical purity of the test substance:

Structural formula: (position of label)

0

No[x]

Mbq/mg (=  $61.37 \,\mu\text{Ci/mg}$ )

No

2.12 Mbq/mg (=57.82 μCi/mg) / 2.25

Yes [x]

Yes [

Formulation used for study:

Type of formulation (if used):

Co-solvent for application (if used): acetonitrile

See 4.1 4.2 Specification

Storage stability 4.3 The chemical is stable when stored at about -20°C in the dark

Stability in vehicle The test substance was found to be stable in the vehicle when analysed by HPLC before

and after application

Homogeneity in vehicle The test substance was prepared as a homogeneous solution in acetonitrile

Validity 4.6

Vehicle / solvent 4.7 Acetonitrile Lichrosolv® (Merck Darmstadt, Germany)

Physical form crystalline, slightly beige 5.1 Test method Richtlinie für die amtliche Prüfung von Pflanzenschutzmitteln, Teil IV, 4-1: "Verbleib von

Pflanzenschutzmitteln im Boden: Abbau, Umwandlung und Metabolismus." Biologische Bundesanstalt für Land und Forstwirtschaft, Bundesrepublik Deutschland, Dezember,

1986.

5.2 Justification not applicable

5.3 Copy of method available on request

6 Choice of method not applicable

7 Deviations none 8.1 Certified laboratory yes

**8.2 Certifying authority** Federal Department of Interior, Switzerland

**8.3 GLP** GLP Switzerland based on OECD

8.4 Justification not applicable
 9.1 GEP Not applicable
 9.2 Type of facility (official or officially recognised)

9.3 Justification Not applicable

## 10 Test system

System		1	2	3	4	5
Origin of soil:		Collombey	Speyer	Weide	Pappelack	Weide
		(stored in	2.1	(stored in	er	
		the	(stored in	the		(field fresh)
		greenhous	the	greenhous	(field fresh)	
		e)	greenhou	e)		
			se)			
Batch-No:		6/95	89	6/95	PA_5/96	WE_5/96
Analysis date:		Nov. 6, 95	May 4,95	July 20, 95	June 5, 96	June 5, 96
Classification (USDA):		loamy	sand	sandy	loamy	sandy
		sand		loam	sand	loam
Particle size distribution:	% silt	15.2	7.6	31.5	22.7	31.8
	% sand	78.6	88.3	62.3	74.4	63.8
	% clay	6.2	4.1	6.2	2.9	4.4
Organic matter content:	(%)		r	ot given in rep	ort	
Organic carbon content:	(%)	1.7	0.6	1.3	1.1	1.3
Total nitrogen:	(%)	0.21	0.18	0.14	0.1	0.14
pH:		7.4	8.2	7.55	7.6	7.5
CaCO <sub>3</sub> :	(%)	7.2	0.1	11	9.7	10.2
Cation exchange capacity:	(meq/ 100g soil)	14.3	6	14.5	6.5	8.2
Bulk density (air dried and sieved (2 mm ) soil)	(g/ml)		r	not given in rep	oort	
Maximum water holding capacity (MWC; pF<0.3):	(ml H <sub>2</sub> O/ 100g dry soil)	46.7	23.5	47.2	39.5	47.4
Field capacity (FC; pF=2.5):	(ml H <sub>2</sub> O/ 100g dry soil)	36.2	18.8	36.3	29.3	36.6
Microbial biomass (mg/100 g	At start:	37.1	11.8	36.7	35.5	29.6
dry soil):	At end:	26.6	6.5	18.0	17.8	22.8
Soil conditions	Aerobic:	Х	х	X	Х	Х
Soil moisture:	%- MWC:	40%	40%	40%	40%	40%

Test conditions		1	2	3	4	5
Incubation temperature(s)	(°C)	20±2	20±2	20±2	20±2	20±2
Treatment rate:	(mg a.i./ kg soil)	0.496	0.496	0.496	0.492	0.492
Incubation time:	(days)	181	181	181	121	121
Number of samples taken for analysis:		24	24	24	10	10
Methods used for analysis	HPLC	Х	х	Х	х	х
The state of the s	TLC	Х	Х	Х	Х	Х
	GC	-		- 1	-	·-
Methods for the identific degradates	cation of		cochro	matography, L	C-MS	

11 Statistics The rate of disappearance of CGA 293343 was calculated by applying first order one- and two- compartment reaction kinetics

12 References (published) none13 Unpublished data none

Findings: Details of the degradation pattern and the radioactivity balance are given in Table 1. The percentage of extractable radioactivity decreased in all soil types with time. The main extractable metabolite observed was CGA 322704 in all soils. It reached maximum amounts corresponding to between 4.6% of applied radioactivity in the Speyer 2.1 soil after 181 days and 18.9% in the Weide field soil after 121 days. CGA 322704 was identified by standard reference cochromatography and by mass spectroscopy. An additional metabolite identified was the hydrolysis product CGA 355190. It was observed in all soils except in the Speyer 2.1 soil, in amounts ≤2.4% of the applied dose. Unknown chromatographic fractions consisted of up to 5 individual compounds reaching in sum 6.5% of the applied dose as a maximum.

Bound residues reached a maximum concentration between 10.2% and 16.6% after 181 days in the greenhouse soils and of 7.6% and 9.6% after 121 days in the field soils. Volatile radioactivity was identified as carbon dioxide. It reached between 12.1% and 21.1% after 181 days in the greenhouse soils and between 12.1% and 14.2% after 121 days in the field soils.

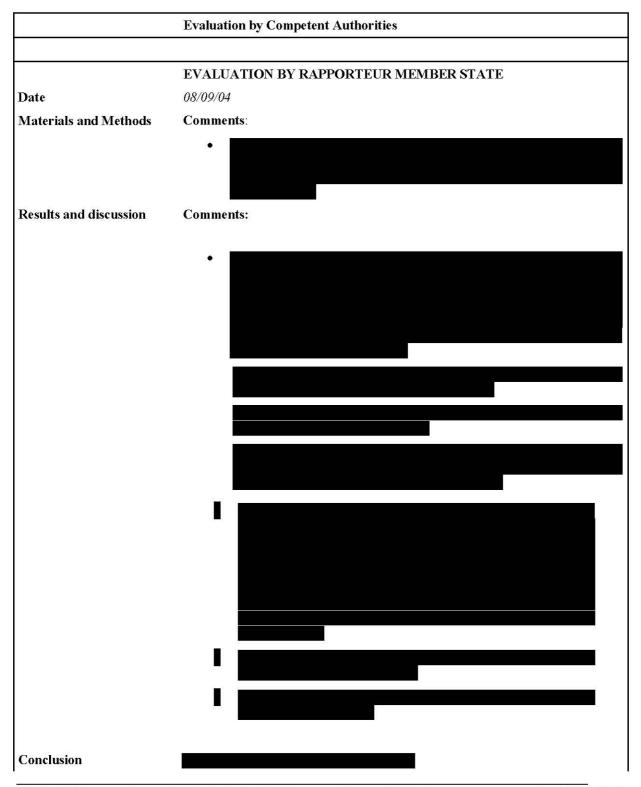
The data show that the route and pattern of the thiamethoxam degradation is independent of the soil type.

Table 1: Distribution and recovery of radioactivity in percent of applied <sup>14</sup>C-thiamethoxam in various soils under laboratory conditions (Adam 1996a).

Days	Volatiles	Extrac-	Extract.	CGA-	CGA-	CGA-	Un-	Non-	Total
after applic.	(CO <sub>2</sub> ) [%]	tables [%]	Soxhlet [%]	293343 [%]	322704 [%]	355190 [%]	known* [%]	extrac- tables [%]	[%]
Collombe	y Soil store	d in areen	house					[/0]	
0	n.d	99.0	n.d.	99.0	< 0.5	< 0.5	< 0.5	2.2	101.2
3	0.3	97.2	2.7	97.7	1.2	< 0.5	1.0	0.6	100.7
7	0.7	95.6	2.9	96.4	1.8	< 0.5	< 0.5	1.2	100.4
14	1.4	92.9	3.5	90.8	1.1	< 0.5	4.5	1.6	99.4
21	1.7	91.0	3.3	88.1	4.5	< 0.5	1.7	2.3	98.3
28	2.3	90.9	3.4	86.5	4.4	< 0.5	3.2	2.3	98.8
45	3.9	87.1	4.3	79.2	7.8	< 0.5	4.4	3.0	98.3
59	5.2	84.5	4.7	77.2	8.2	< 0.5	3.5 4.4	4.4	98.8
90	7.7	79.2	4.4	67.5	11.3	< 0.5		5.8 7.0	97.1
120	9.5 11.7	76.1 71.4	5.0 5.3	61.9 54.8	12.4 14.3	< 0.5	6.5 6.3	7.0 8.0	97.5 96.5
153 181	13.5	68.2	5.5 5.6	54.6 53.7	12.9	1.3 1.2	5.4	10.2	96.5
	Soil store			33.7	12.8	1.2	3.4	10.2	91.4
0	n.d.	101.2	n.d.	101.2	< 0.5	< 0.5	< 0.5	1.1	102.3
3	0.4	96.3	n.d.	95.2	0.3	< 0.5	0.8	3.7	100.4
7	0.8	96.5	2.6	96.7	0.6	< 0.5	1.7	2.9	102.8
14	1.3	89.1	4.1	87.1	3.2	< 0.5	2.9	4.2	98.7
21	1.9	90.3	4.4	89.0	1.4	< 0.5	4.3	4.9	101.6
28	2.6	87.2	5.5	87.8	1.7	< 0.5	3.2	5.2	100.4
45	3.9	84.1	7.0	83.6	2.3	< 0.5	5.1	5.8	100.8
59	5.0	78.9	6.2	77.9	2.5	< 0.5	4.6	9.0	99.0
90	7.1	76.3	5.6	73.6	3.0	< 0.5	5.2	11.6	100.6
120	9.0	72.1	8.5	70.7	3.9	< 0.5	6.0	11.9	101.4
153	10.8	69.0	6.9	66.6	4.2	< 0.5 < 0.5	5.2 5.4	14.1	100.9
181	12.1 I stored in	63.7	7.4	61.1	4.6	< 0.5	5.4	16.6	99.8
0	n.d.	97.8	n.d.	97.6	0.3	< 0.5	< 0.5	2.5	100.4
3	0.5	94.9	2.6	95.8	1.3	< 0.5	0.3	0.7	98.6
7	0.6	95.5	3.3	97.1	1.2	< 0.5	0.5	1.3	100.7
14	1.7	90.3	4.1	88.0	3.2	< 0.5	3.3	1.8	97.9
21		90.0	3.4	86.9	4.2	< 0.5	2.3	2.9	99.1
28	2.8 3.2	86.2	4.2	82.3	5.2	< 0.5	2.9	3.0	96.6
45	5.8	84.0	5.1	77.4	7.2	< 0.5	4.5	3.9	98.8
59	7.9	79.6	5.2	71.2	9.2	< 0.5	4.5	5.9	98.6
90	12.0	72.5	5.3	59.4	12.3	< 0.5	6.1	7.5	97.3
120	15.3	68.8	6.0	55.8	13.0	< 0.5	6.0	8.8	99.0
153	18.6	63.5	5.6	47.5	14.6	0.8	6.2	10.0	97.8
181	21.1	63.8	5.6	47.0	15.1	1.2	6.1	12.5	102.9
Weide fiel	·······	OE E	0.0	07.0	- O E	0.0	- O E	0.1	07.0
0	n.d. 0.6	95.5 95.3	2.3 3.3	97.0 96.3	< 0.5 1.6	0.9 0.6	< 0.5	0.1 0.7	97.9 99.9
3 7	1.4	91.7	3.6	90.3	2.6	0.6	< 0.5 0.7	0.7 1.2	99.9 97.8
14	2.8	88.6	4.1	86.4	4.8	0.8	0.7	1.9	97.4
21	4.4	89.4	3.5	80.6	8.0	1.0	3.3	2.5	99.8
28	5.2	82.0	4.5	72.8	9.5	0.8	3.4	3.7	95.4
45	8.0	79.1	4.8	66.2	12.8	< 0.5	4.8	4. 9	96.8
62	10. 5	73.3	3.4	55.1	14.2	1.2	6.2	6.3	93.4
90	13.3	64.8	4.9	44.6	16.8	0.8	7.5	8.5	91.5
121	14.2	56.9	6.0	36.3	18.9	< 0.5	7.6	9. 6	86.6
	cer field soi							,	
0	n.d.	96.6	2.0	97.3	< 0.5	1.4	< 0.5	0.1	98.7
3 7	0.4	96.2	2.9	96.8	1.0	1.2	< 0.5	0.8	100.2
	1.0	90.3	4.1	91.5	1.6	1.3	< 0.5	1.2	96.6
14	2.2	87.8	4.3	87.5	3.0	1.6	< 0.5	1.6	95.9
21	3.4	89.2	3.6	86.6	5.1	< 0.5	1.0	2.4	98.4

Days after applic.	Volatiles (CO₂) [%]	Extrac- tables [%]	Extract. Soxhlet [%]	CGA- 293343 [%]	CGA- 322704 [%]	CGA- 355190 [%]	Un- known* [%]	Non- extrac- tables [%]	Total [%]
28	4.4	81.7	6.0	78.7	5.8	2.4	0.8	3.9	96.0
45	6.6	79.6	5.5	73.4	8.3	1.0	2.4	4.1	95.7
62	8.6	77.9	3.8	68.0	10.1	1.3	2.3	5.1	95.3
90	10.9	69.3	6.7	60.8	11.8	1.0	2.4	7.7	94.6
121	12.1	63.3	7.8	54.7	12.9	< 0.5	3.6	7.6	90.8

<sup>\*:</sup> Unknown chromatographic fractions (up to 5 compounds).



Remarks

98/8 Doo section No.		7.2.2.1 <i>I</i> 02a	The rate and route of fegradation including identification of the process involved and identification of any metabolites and degradation products in at least three soil types under appropriate conditions
91/414	Annex	11	Rate of degradation in soil - laboratory studies
Point addre	ssed	7.1.1.2.1 <i>I</i> 02	

1.	Annex point(s)	IIA,7.1.1.2.1 Rate of degradation in soil - laboratory studies						
2.	Location in Dossier	Section 5						
3.	Authors (year) Title	Ellgehausen, H. (1998a) Calculation of the Degradation Kinetics of Metabolite CGA 322704 in Sandy Soil Collombey						
	Report No., Date Source Owner Syngenta File No. (DESIRE)	Report No 98EH04, October 9, 1998 Environmental safety/ Ecochemistry, Novartis Crop Protection AG, Basel, Switzerland Syngenta Crop Protection AG, Basel, Switzerland CGA 322704/16						
4.	Testing facility	Environmental safety/ Ecochemistry, Novartis Crop Protection AG, Basel, Switzerland						
5.	Dates of work	September - October 9, 1998						
6.	Test substance	not applicable						
7.	Test method	The study was performed to generate supplemental data to satisfy the following guidelines:  Commission Directive 95/36/EC amending Council Directive 91/414/EEC, Annex 1; fate and Behaviour in the Environment: 7.1 Fate and Behaviour in Soil, 7.1.1.2 rate of degradation laboratory studies, 1997.						
8.	Deviations	none						
9.	GLP	not applicable						

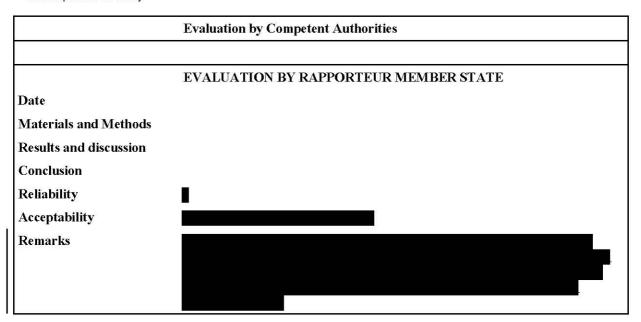
Findings: The rate of disappearance of thiamethoxam was calculated by applying first order one - and two - compartment reaction kinetics. The values are presented in Table 1. The data indicate that degradation in soils taken directly from the field is somewhat faster as compared to soils stored before use. The comparably long half-life of 272 days in the Speyer 2.1 soil with low microbial activity indicates that thiamethoxam degradation is a microbially driven process.

The data in the Collombey sandy soil allowed the calculation of kinetics of the dissipation of the metabolite CGA 322704 by using the ModelMaker software (Version 3.0.2; SB Technology Ltd.). The DT<sub>50</sub> obtained was 75 days with a corresponding DT<sub>90</sub> of 250 days. In the other soils tested CGA 322704 did not reach an apparent peak concentration during the experimental period of 121 and 181 days, respectively.

Table 1: Results of degradation kinetics for thiamethoxam and amounts of bound residues and carbon dioxide in five different soil systems (Adam 1996a)

Soil	Incubation time	thiamethoxam	thiamethoxam	$^{14}\mathrm{CO}_2$	Non extractable
	[days]	DT <sub>50</sub> [days]	DT 90 [days]	[% appl. act.]*	[% appl. act.]*
Collombey (greenhouse)	181	184	694	13.5	10.2
Speyer 2.1 (greenhouse)	181	272	1013	12.1	16.6
Weide (greenhouse)	181	143	475	21.1	12.5
Pappelacker (field)	121	164	911	12.1	7.6
Weide (field)	121	80	266	14.2	9.6

<sup>\*</sup> at completion of study



98/8 Doc IIIA section No.	7.2.2.1 <i>I</i> 03	The rate and route of degradation including identification of the process involved and identification of any metabolites and degradation products in at least three soil types under appropriate conditions
91/414 Annex Point addressed	II 7.1.1.1.1 / 02	Soil route of degradation: Aerobic degradation

1. Annex point(s) II A, 7.1.1.1.1 Soil Route of Degradation: Aerobic degradation 2. Location in Dossier Section 5. 3. Authors (year) B.Dixon (1998) Aerobic Soil Metabolism of (14C-thiazole) CGA 293343 Title Report No., Date ABR-96059, March 16, 1998 Syngenta File N°(Desire) 293343/478 Syngenta Crop Protection AG Owner **Testing facility** Novartis Crop Protection Inc Environmental Safety Department Greensboro, NC 27419, USA PTRL West, Incorporated Richmond, CA 94806, USA 5. Dates of work Study Initiation: 25 September, 1995 Experimental Start: 16 October, 1995 Experimental Termination: Study Completion: 16 March, 1998 6. Test substance ISO common name thiamethoxam. Company Code: CGA 293343, (14C-thiazolyl)CGA 293343, Batch specific radioactivity: 1.09 MBq/mg radiochemical purity: 7. Test method US EPA Environmental Fate Data requirement, 40 CFR Section 158, Subdivision N, series 162-1. 8. Deviations none Yes, EPA Good Laboratory Practice Standards (40 CFR Part 160) 9. GLP with the exception of the soil characterization performed at Agvise Laboratories Inc.

Test system: In this study the degradation of <sup>14</sup>C-thiazol-labelled thiamethoxam was investigated in a Californian sandy loam soil for one year. The biological portion of the study was contracted at PTRL West, Inc., Richmond, California (study No. 573W). Soil was dosed at two concentrations, 0.09 mg/kg (kinetic viable test, corresponding to a field rate of 200 g/ha assuming a homogeneous distribution in the top 15 cm soil layer and a soil density of 1.5 g/cm<sup>3</sup>) and 4.98 mg/kg (bulk viable test). Incubation was at 25°C and 75% field moisture capacity in the dark under aerobic conditions for 365 days. Soil samples were aerated and trapped for volatiles continuously. Duplicate samples were taken at regular intervals (12 for the 0.09 mg/kg and 8 for the 4.98 mg/kg incubations) to determine the metabolism occurring. Additional tests were performed under initially sterile conditions. Details of the soil characteristics are summarised in Table 1.

Table 1: Parameters of soil used for thiamethoxam metabolism (Dixon, 1998 and Schwartz, 1998a)

Location of collect	ion	Novartis Crop Pro	ntection \Mestern			
Location of conco			Novartis Crop Protection, Western Research Station, Sanger, California			
Date of collection (f	ield/greenhouse)	September				
На			3			
Organic matter (%)		l o	6			
CEC (meg/100 g so	il)	7.	.4			
Water holding capa		10	10.9			
Classification (USD		sandy	sandy loam			
Particle size: `	Clay (%)		3			
	Silt (%)	2	5			
	Sand (%)	6	7			
Microbial biomass (	mg C/100 g soil):	14C-Thiazol study	<sup>14</sup> C-Guanidine			
			study			
day 0		17.5	17.5			
4.5 months		29.0	20.0			
12 months		19.0	23.5			

Findings: For both the 0.09 and the 4.98 mg/kg incubations the amount of extractable radioactivity decreased steadily during the 365 days of incubation. Results for the dissipation of thiamethoxam and the formation and decline of metabolites, bound residues and carbon dioxide are shown in Table 2 and Figure 1 for the 0.09 mg/kg concentration (kinetic viable test). A quantitatively and qualitatively similar pattern was observed for the 4.98 mg/kg incubation (bulk viable test).

For the kinetic viable test the sum of the extracts 1 and 2 (acetonitrile and acetonitrile-ammonium chloride solution at room temperature) decreased from 99.7% on day 0 to 52.0% at day 365. Most of the extractable radioactivity in the extracts was identified as undegraded thiamethoxam. There were multiple minor components detected by two dimensional TLC. Some of these minor components were observed only intermittently during the 365 days. Components characterised by co-chromatography were CGA 322704 and CGA 355190 at concentrations below 5% of applied dose and CGA 309335 and CGA 353968 at concentrations below 1.5% of applied dose. With the exception of CGA 309335 all other components identified had also been observed in the soil metabolism study with the <sup>14</sup>C-guanidine-labelled material.

Aerobic volatile generation of carbon dioxide increased to an average of 11% at day 268 and to 7% at day 365 indicating mineralisation of the thiazol ring. The non-extractables reached a maximum of 28% at day 365. Harsh extraction under strong alkaline conditions (1M NaOH) had released up to 18% of applied radioactivity mostly in the fulvic acid fraction.

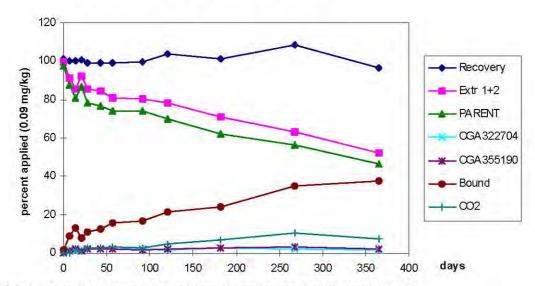
The initial degradation of thiamethoxam in the sterile soil was mainly due to hydrolytic degradation. CGA 355190 was one of the major degradates in the hydrolysis studies. This degradate accumulated under the initially sterile conditions whereas in the viable soil incubations it was also formed as an intermediate but was metabolized further.

Table 2: Distribution and recovery of radioactivity after application of <sup>14</sup>C-thiamethoxam in a Californian sandy loam, kinetic viable incubation (Dixon 1998).

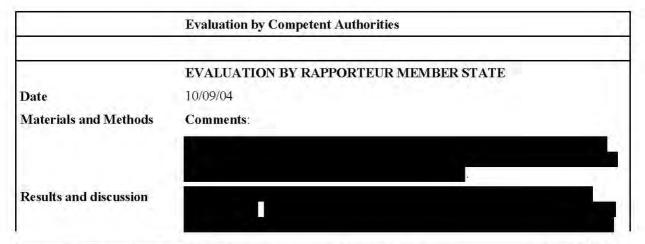
Time after applic. [days]	Volatiles (CO <sub>2</sub> ) [%]	Extract. 1 + 2 [%]	CGA- 293343 [%]	CGA- 322704 [%]	CGA- 355190 [%]	Harsh Extract [%]	Non- extrac- tables [%]	Total Recovery [%]
0	np	99.7	97.4	0.9	0.2	np	1.4	101.1
7	0.3	91.1	87.5	1.3	1.0	np	8. 7	100.0
14	1.7	85.4	81.1	1.1	2.1	np	13.3	100.4
21	1.0	92.1	86.6	1.6	1.0	np	7.7	100.8
28	2.6	85.6	78.1	1.9	2.2	7.3	3.9	99.4
43	2.4	84.5	76.5	1.9	2.2	np	12.3	99.2
57	3.1	80.7	73.9	2.1	2.3	np	15.6	99.4
92	2.6	80.4	73.9	1.9	1.8	9.3	7.6	99.9
121	4.8	78.0	69.9	1.7	2.3	np	15.2	98.0
182	6.7	70.7	61.9	2.4	2.5	11.5	9.6	98.2
268	10.7	62.9	56.2	2.1	3.0	17.7	17.4	108.6
365	7.1	52.0	46.2	1.5	2.3	9.2	28.1	96.6

np: not performed

Figure 1: Formation and decline of major metabolites for <sup>14</sup>C-Thiazol-thiamethoxam in a Californian sandy loam (Dixon 1998)



Bound = sum of non-extractable radioactivity and radioactivity in harsh extracts





Section 7: Ecotoxicological Profile Including Environmental Fate and Behaviour

Reliability indicator

Remarks

98/8 Doc IIIA section No.	7.2.2.1 <i>I</i> 04	The rate and route of degradation including identification of the process involved and identification of any metabolites and degradation products in at least three soil types under appropriate conditions
91/414 Annex Point addressed	II 7.1.1.1.1 / 03	Soil route of degradation: Aerobic degradation

1. Annex point(s) II A, 7.1.1.1.1 Soil Route of Degradation: Aerobic degradation 2. Location in Dossier Section 5. 3. Authors (year) B.Schwartz, (1998a) Title Final report: Aerobic Soil Metabolism of (14C-guanidine) CGA 293343 Report No., Date ABR-96084, March 3, 1998 Syngenta File N°(Desire) 293343/453 Syngenta Crop Protection AG Owner 4 **Testing facility** Novartis Crop Protection Inc **Environmental Safety Department** Greensboro, NC 27419, USA PTRL West, Incorporated Richmond, CA 94806, USA Dates of work Study Initiation: 25 September, 1995 5. Experimental Start: October 3, 1995 Experimental Termination: Study Completion: 16 March, 1998 6. Test substance ISO common name thiamethoxam. Company Code; CGA 293343, (14C-guanidine)CGA 293343. specific radioactivity: 1.57 MBq/mg radiochemical purity: US EPA Environmental Fate Data requirement, 40 CFR Section 158, Subdivision N, 7. Test method Series 162-1. Deviations 8. none 9. Yes, EPA Good Laboratory Practice Standards (40 CFR Part 160) with the exception of the soil characterization performed at Agvise Laboratories Inc.

Test system: In this second aerobic soil metabolism study the degradation of thiamethoxam was investigated by using <sup>14</sup>C-guanidine-labelled material. As above the biological portion of the study was contracted at PTRL West Inc., Richmond, California (study No. 572W). The dosage levels were 0.09 mg/kg (kinetic viable test, corresponding to a field rate of 200 g/ha assuming a homogeneous distribution in the top 15 cm soil layer and a soil density of 1.5 g/cm³) and 4.90 mg/kg (bulk viable test). Test soil, incubation conditions and soil sampling regimen were the same as for the study with the <sup>14</sup>C-thiazol-labelled material. Additional tests were performed under initially sterile conditions.

Findings: As with the <sup>14</sup>C-thiazol-labelled material the amount of extractable radioactivity decreased steadily during the 365 days incubation for both concentrations. Results for the dissipation of thiamethoxam and the formation and decline of metabolites, bound residues and carbon dioxide are shown in Table 1 and Figure 1 for the 0.09 mg/kg concentration (kinetic viable test). A quantitatively and qualitatively similar pattern was observed for the 4.98 mg/kg incubation (bulk viable test).

The sum of the extracts 1 and 2 (acetonitrile and acetonitrile-ammonium chloride solution at room temperature) decreased from 100% on day 0 to 53% at day 365. Most of the extractable radioactivity in the extracts was identified as undegraded thiamethoxam. There were up to 30 multiple minor components detected by two dimensional TLC. Some of these minor